Adsorption Analysis of Fluoride Removal Using Graphene Oxide/Eggshell Adsorbent

Norhusna Mohamad Nor¹, Nur Hidayatul Nazrah Kamil¹, Amirul Izan Mansor¹, and Hawaiah Imam Maarof²

¹Faculty of Chemical Engineering, Universiti Teknologi MARA, Cawangan Pulau Pinang, 13500 Permatang Pauh, Pulau Pinang, Malaysia
²Faculty of Chemical Engineering, Universiti Teknologi MARA, 40450 Shah Alam, Selangor, Malaysia

*Corresponding author:
tel: +604-3822539
e-mail: norhusna8711@ppinang.uitm.edu.my
Received: February 8, 2019
Accepted: July 12, 2019
DOI: 10.22146/ijc.43481

Abstract: Graphene oxide with eggshells (GO/ES) adsorbent has been studied for fluoride ions (F⁻) removal. An adsorption study was conducted in batch experiments at different adsorption parameters, which are initial F⁻ concentration, contact time, and temperature. The effects of these adsorption parameters towards F⁻ removal by using GO/ES adsorbent were investigated. The adsorption parameters were then analyzed with adsorption isotherms (Langmuir and Freundlich), kinetics (pseudo-first-order and second-order) and thermodynamic studies. Under various parameters, GO/ES is proven as an effective adsorbent with an adsorption capacity of F⁻ up to 48 mg/g. The experimental data were satisfactorily fitted with Langmuir isotherm, which illustrated the monolayer pattern of F⁻ adsorption into GO/ES adsorbent. The adsorption kinetic analysis indicated that the adsorption data could be well described by Pseudo-second-order kinetic model, which indicated the chemisorption process, while thermodynamic studies revealed that the adsorption of F⁻ was an exothermic process.

Keywords: graphene oxide; eggshell; fluoride; adsorption; isotherms; thermodynamics; kinetics

INTRODUCTION

Contamination of fluoride (F⁻) in groundwater has been identified as a serious problem, which contributes to adverse effects on human health, aquatic life and indirectly to the environment as well. This occurrence is contributed by various sources, such as effluent from industrial activities, agricultural activities, and existing geological formation [1]. World Health Organization (WHO) has classified F⁻ as one of the main contaminants in water; hence, a maximum allowable limit for F⁻ in drinking water is suggested to be as low as 1.5 mg/L [3]. In addition, the Ministry of Health Malaysia has also set regulation for the National Standard for Drinking Water Quality, where it must be below than 0.4–0.6 ppm of F⁻ concentration [4]. According to the Environmental Quality (Industrial Effluent) regulations, the permissible F⁻ limit for industrial effluent in standard A and B are 2.0 mg/L and 5.0 mg/L, respectively [5]. Therefore, the wastewater effluent discharged from all industry must meet the allowable limit. Hence, concerning human health, developing efficient methods in F⁻ removal is significant to ensure F⁻ concentration is within acceptable limits.

The removal of F⁻ has become a great concern globally due to its toxic effect on the water towards humans and the environment. Various treatment methods are available for F⁻ removal, for instance, ion exchange, chemical precipitation, fluidized-bed precipitation, aluminum-solublity, membrane filtration, electrochemical method, and adsorption [1-2]. Among those methods, the adsorption process has been identified as the most efficient and economical method to be used in F⁻ removal [6]. The efficiency of adsorption depends on the type of adsorbents used, where high removal of F⁻ is anticipated based on the excellent development of the adsorptive materials. There are
many types of adsorbents that have been used for fluoride removal, such as activated carbon, biochar, raw biomass, fly ash, calcium-containing materials, and graphene oxide (GO) [1].

Calcium containing materials such as bone, eggshells, calcium carbonate (CaCO₃), and seashells have been used by researchers as adsorbents for F⁻ removal, due to their strong affinity towards the target pollutant [7]. GO is the combination of graphene and oxidizing agents, producing hydrophilic materials, and expands layer of separation [8]. The unique properties of GO are, it is a promising material for adsorption of F⁻, due to its structure which can be altered, and it has excellent chemical, thermal and mechanical stability [8-9]. Furthermore, it also offers high surface areas with abundant binding sites, which is suitable for further modification. In this research work, the modification of GO with calcium-containing material, which is eggshells (ES) will be discussed in detail. ES has been selected, because of its low-cost material (waste) which is abundantly available, and its composition consists of a high percentage of CaCO₃. Thus, its combination with GO is a perfect match for F⁻ removal. The objective of this work is to investigate the effects of various adsorption parameters and adsorption analysis of F⁻ removal by using GO/ES adsorbent.

EXPERIMENTAL SECTION

Materials

Eggshells (ES) were obtained from a nearby cafeteria, UiTM Cawangan Pulau Pinang. Chemicals used for the synthesis of graphene oxide (GO) were graphite powder (99.99%), sodium nitrate (NaNO₃ 98%), sulphuric acid (H₂SO₄ 98%), potassium permanganate, (KMnO₄), hydrogen peroxide (H₂O₂), and hydrochloric acid (HCl). As for the batch adsorption experiment, the chemical used was sodium fluoride (NaF).

Procedure

Synthesis of GO/ES adsorbent

Graphene oxide (GO) was prepared by using a modified Hummer’s method [10]. 2.5 g of graphite powder and 1.25 g of NaNO₃ were added together with 60 mL of H₂SO₄ in a conical flask with continuous stirring for 2 h at constant temperature (10 °C). KMnO₄ (7.5 g) was added into the mixture, where the reaction temperature was controlled to be lower than 10 °C. The mixture was continued stirred at 30 °C for 2 h until the solution becomes pasty brownish. Next, 135 mL of deionized (DI) water was added slowly into the mixture. The reaction temperature was increased to 50 °C until the brownish color obtained. Finally, the mixture was terminated by the addition of 25 mL of H₂O₂, which turned the mixture into gold-yellow color. The mixture was filtered and washed with 400 mL of HCl and DI water repeatedly to remove any impurities. Then, it was dried in a drying oven at 60 °C for 24 h [11].

Pre-treatment of ES was done by washing it with DI water and drying in an oven at 45 °C for overnight. Then, ES was ground into a powder form, to obtain 250 μm size of particles size. One gram of GO in a slurry form was added into a beaker containing 300 mL of DI water and sonicated for 1 h at 45 °C. Then, 5 g of ES was added into the mixture. After 1 h, the suspension was centrifuged for 15 min at 3000 rpm and dried at 60 °C overnight [12].

Batch experimental of fluoride removal

The effects of adsorption parameters (i.e., F⁻ initial concentration, contact time, and temperature) were conducted via a batch experiment. A stock solution of 1000 mg/L of fluoride was prepared by dissolving 1 g of NaF in 1 L of distilled water. The stock solution was diluted with distilled water to an achieved a concentration of 30 mg/L. A 200 mL of NaF stock solution was tested with 0.05 g of GO/ES adsorbent at 25 °C. The mixture was shaken by using an orbital shaker with 200 rpm for 60 min. The adsorbent was filtered through a 0.45 μm membrane filter. The steps were repeated at different concentrations (30, 40, 50, 60, 70 mg/L) contact times (30, 60, 90, 120 min) and temperatures (25, 45, 60, 75 °C). F⁻ final concentrations were measured by using DR 2800 Spectrophotometer before and after adsorption. SPADNS reagent was used as a standard to detect F⁻ concentration in samples. The adsorption capacity and percentage removal (%R) of F⁻ were calculated by using Eq. (1) and (2), respectively.
where \( Q_e \) is F\( \text{–} \) adsorbed per unit mass of adsorbent (mg/g), \( C_t \) is the final concentration of F\( \text{–} \) at equilibrium (mg/L), \( C_0 \) is an initial concentration of F\( \text{–} \) (mg/L) and \( m \) is mass of adsorbent (g/L).

### Adsorption isotherms

The analysis of adsorption isotherms were linearized as in Eq. (3) and (4) that represent Langmuir and Freundlich isotherms, respectively.

\[
\frac{C_e}{q_e} = \frac{C_e}{q_{\text{max}}} + \frac{1}{q_{\text{max}}K_L}
\]

\[
\log q_e = \log K_F + \frac{1}{n}\log C_e
\]

where \( C_e \) is equilibrium concentration of fluoride (mg/L), \( q_e \) is amount of fluoride adsorbed at equilibrium per gram of adsorbent (mg/g), \( q_{\text{max}} \) is adsorption for a complete monolayer (mg/g), \( k_L \) is a constant value of Langmuir (L/mg), \( K_F \) is Freundlich constant (L/mg), and \( 1/n \) is adsorption intensity.

### Adsorption kinetics

The pseudo-first-order and pseudo-second-order adsorption kinetic models were used to investigate the adsorption kinetics of fluoride and to quantify the extent of uptake in the adsorption process, as shown in Eq. (5) and (6), respectively.

\[
\log(q_e - q) = \log q_e - \frac{K_1t}{2.303}
\]

\[
\frac{t}{q} = \frac{1}{K_2q_e^2} + \frac{1}{q_e}\frac{t}{1}
\]

where \( q_e \) is equilibrium adsorption capacity (mg/g), \( q \) is adsorption capacity at times (mg/g), \( K_1 \) is rate constant of pseudo-first-order adsorption, \( K_2 \) is rate constant of pseudo-first-order adsorption, and \( t \) is time.

### Adsorption thermodynamics

Thermodynamic parameters such as standard free energy change (\( \Delta G^\circ \)), enthalpy change (\( \Delta H^\circ \)) and entropy change (\( \Delta S^\circ \)) were calculated to evaluate the spontaneous nature and the thermodynamic feasibility of the adsorption process [13]. The value of the equilibrium constant, \( K_0 \), is calculated by using Eq. (7). The dependence of temperature on the adsorption of fluoride ions onto GO/ES adsorbent was evaluated using Eq. (8) and (9).

\[
K_0 = \frac{C_0}{C_e}
\]

\[
\Delta G^\circ = -RT\ln K_0
\]

\[
\log K_0 = \frac{\Delta S^\circ}{2.303R} - \frac{\Delta H^\circ}{2.303RT}
\]

where \( C_0 \) is the initial amount of adsorbate in solution (mg/L), \( C_e \) is equilibrium amount of adsorbate in solution (mg/L) \( R \) is universal gas constant (J/mol.K), and \( T \) is temperature (K).

### RESULTS AND DISCUSSION

Batch adsorption studies were conducted on various adsorption parameters, which are F\( \text{–} \) initial concentration, temperature, and contact time. The results of adsorption capacity and % removal of F\( \text{–} \) at different adsorption parameters are shown in Fig. 1, respectively. Details discussion on the effect of each parameter will be discussed in the following subsections.

According to Kashi et al. the eggshell is a promising material to be used as an adsorbent for F\( \text{–} \) removal, which the fluoride removal was up to 100% [7]. The combination of eggshell and GO is expected to produce an excellent adsorbent since the use of GO modified with magnetic iron and aluminum oxide done by Liu et al. could remove F\( \text{–} \) up to 64 mg/g [11]. The combination of GO with eggshells (GO/ES) in this research work can adsorb up until 56 mg/g.

### Effect of Initial Fluoride Concentration

To study the effect of initial F\( \text{–} \) concentration, the experiments were carried out with 0.05 g of adsorbent dosage, 60 min of contact time, and 25 °C of temperature. The % removal of F\( \text{–} \) as shown in Fig. 1(a) decreases as F\( \text{–} \) initial concentration increases. However, the adsorption capacities of F\( \text{–} \) shown in Fig. 1 do not vary significantly, which are in the range of 44–46 mg/g. This is due to the vacant sites in GO/ES adsorbent which are capable of adsorbing up to a certain capacity, even though the higher initial concentration of F\( \text{–} \) was introduced. The ratio of F\( \text{–} \) to the number of available sites in GO/ES adsorbent is constant, while the vacant sites were saturated with F\( \text{–} \).
at certain concentration [7,13]. As the concentration gradient of $F^-$ increased, the driving force will increase to overcome the mass transfer resistance of $F^-$ from aqueous solution into a solid phase, which increases the adsorption equilibrium of GO/ES adsorbent until it achieved a saturated state. At the same time, this occurrence also resulted from the fixed adsorbent dosage used, where available vacant sites are limit with the increase of $F^-$ initial concentration. These results are in agreement with the study done by Bhaumik et al. [13], where an increase of $F^-$ concentration resulted in the decrement of the % removal.

**Effect of Contact Time**

The effect of contact time was varied at a given initial concentration (25 mg/L), adsorbent dosage (0.05 g) and temperature (25 °C). As illustrated in Fig. 1(b), the % removal of $F^-$ increased as contact time increases from 30 to 90 min and remained constant afterward, where no significant changes are observed from 90 to 120 min. A similar trend is shown by the adsorption capacity of $F^-$ in Fig. 1, where the values increase from 36 to 55 mg/g and start to reach equilibrium from 55–56 mg/g. The modification of ES into GO resulted in a large number of active binding sites, where the positive charge on Ca$^{2+}$ extracted from CaCO$_3$ may be existed on GO’s surface, thus enhancing the adsorption rate of $F^-$ that is having negative charge [7]. Increasing the availability of active binding sites on the adsorbent surface might result in a high removal of $F^-$. In this work, the adsorption of $F^-$ to reach equilibrium is started after 90 min, which is slower compared to other previous works [13-14]. It is presumed that the diffusion process of $F^-$ from the bulk solution to the GO/ES adsorbent’s surface is controlled by the affinity of $F^-$ towards GO/ES’s active binding sites, where longer time is needed for adsorption to reach equilibrium.

**Effect of Temperature**

The adsorption temperature of $F^-$ was varied at 25, 45, 60 and 75 °C, with 25 mg/L of $F^-$ initial concentration, 0.05 g of adsorbent dosage and 60 min of contact time. Fig. 1(c) showed that the % removal of $F^-$ decreased with increase in temperature. This indicates that low temperature (25 °C) is favorable for $F^-$ adsorption, which the highest % removal and adsorption capacity obtained were 48.12% and 48.12 mg/g, respectively. Inefficient removal of $F^-$ was observed at elevated temperature, starting from 60 to 75 °C. Possibly, this is because high energy was received by $F^-$, resulting in exciting movement of $F^-$ molecules. Hence, the tendency of the molecules to escape from active binding sites is greater, which lead to less $F^-$ being adsorbed into GO/ES adsorbent as the temperature is increased [15]. Therefore, lower temperature (less than 60 °C) is anticipated to be used to achieve high removal of $F^-$, since sufficient energy will be delivered to the molecules. Indirectly, $F^-$ will be steadily diffused into GO/ES’s active binding sites without any resistances from the exciting movement of the molecules.

![Fig 1. Effect of (a) initial concentration (b) contact time and (c) temperature on $F^-$ percentage removal by using GO/ES adsorbent](image)
Adsorption Isotherms

In this study, the adsorption mechanism of F⁻ onto GO/ES adsorbent is further explained by using isotherm models, which are Langmuir and Freundlich. These isotherms are capable to describe the interaction between F⁻ adsorbate and GO/ES adsorbent, which include homogeneity of the adsorbent and type of adsorption coverage [9]. The results of the isotherms are illustrated in Fig. 2, where higher R² value was obtained via the Langmuir isotherm model (0.9983) compared to the Freundlich isotherm model (0.5007). The experimental data were well fitted with the Langmuir model, where the adsorption of F⁻ is defined as monolayer adsorption onto GO/ES adsorbent’s surface that is containing a certain number of identical sites [8]. Thus, this indicates, the adsorption of F⁻ is not effective for multilayer adsorption onto the heterogeneous surface (Freundlich model).

Constant parameters of adsorption isotherms that are shown in Table 1 were calculated according to the linearized isotherm models. The empirical constants for Langmuir isotherm, which are K_L and q_m were found to be 0.6221 L/mg and 46.73 mg/g, respectively. The K_L constant measured in this study is relatively larger, which indicates a strong affinity between F⁻ (adsorbate) and GO/ES adsorbent [14]. For Freundlich isotherm, the n constant should be obtained in the range of 1–10, to ensure favorable adsorption [13]. However, the n value in this study is greater than 10, which agrees with the low R² value, where the adsorption mechanism of F⁻ onto GO/ES adsorbent is not suitable to represent by Freundlich isotherm. Overall, Langmuir isotherm was very well fitted with this research data, which suggests that the binding energy on the whole surface of the GO/ES adsorbent was uniform and the adsorbed F⁻ were adsorbed by forming monolayer adsorption, where sorbate-sorbate interaction can be neglected [13].

Adsorption Kinetics

The kinetics study of F⁻ adsorption by using GO/ES adsorbent was analyzed by using linearized pseudo-first-order and pseudo-second-order models expressed as in Eq. (5) and (6), respectively with regard to adsorption on the solid surface system [16]. The experimental data obtained were fitted into the models, as illustrated in Fig. 3 and the K values were calculated and tabulated in Table 2.

From Fig. 3, the pseudo-second-order model represents the best F⁻ adsorption kinetic on GO/ES adsorbent compared to pseudo-first-order, with an R².

| Table 1. Parameters of adsorption isotherms constant |
|-----------------------------------|-----|-----|
| Langmuir                          | Freundlich |
| q_m (mg/g)                        | 46.73 | n  | 32.79 |
| K_L (L/mg)                        | 0.6221 | K_r | 0.4721 |
| R²                                | 0.9983 | R² | 0.5007 |

| Table 2. Kinetics parameter for the adsorption of fluoride by GO/ES adsorbent |
|-----------------------------------|-----|-----|
|                                    | Pseudo-first-order | Pseudo-second-order |
| Rate constant (K_1)               | 0.03020          | Rate constant (K_2) | 0.00047          |
| R²                                | 0.95880          | R²                   | 0.96970          |
Fig 3. (a) Pseudo-first-order and (b) pseudo-second-order adsorption kinetics models for F– removal by using GO/ES adsorbent

value of 0.9697 and K2 value of 0.00047. According to pseudo-second-order theory, the adsorption of fluoride onto GO/ES adsorbent is a chemisorption process which involves sharing or exchange of electrons between the adsorbent and F– [13]. There is a possibility that F– may diffuse across a liquid phase prior to GO/ES’s active sites through intraparticle diffusion and indirectly reacts with available active sites, such Ca2+ to produce CaF [17-20].

Adsorption Thermodynamics

Thermodynamic parameters, such as a change in Gibbs free energy (ΔG°), enthalpy (ΔH°) and entropy (ΔS°), were evaluated in this study. K0, which is the equilibrium adsorption constant calculated as the ratio between adsorption concentration and equilibrium concentration of the solution is shown in Eq. (7). The Gibbs free energy (ΔG°) is calculated for adsorption of fluoride onto GO/ES adsorbent at all temperatures, where R is the gas constant (8.314 J/mol.K), and T is the temperature in Kelvin (K). The experimental data were fitted into the linearized Eq. (9), and illustrated in Fig. 4. All thermodynamic parameters were calculated and tabulated in Table 3.

From Fig. 4, the R2 value obtained is 0.9726, which is not as high as the previous study done by other researchers [13]. Perhaps, this is due to less adsorption that happened at high temperature (348.15 K). From Table 3, the positive value of ΔG° indicates that the process is a physic-sorption process. The values of ΔG° in between 0 and to 20 kJ/mol indicate a physic-sorption process, while the values in between -80 to -400 kJ/mol correspond to the chemisorption process [15]. The increment in values of ΔG° with an increase in temperature suggests that higher temperature makes the

![Graph showing Thermodynamics study of F– adsorption onto GO/ES adsorbent](image)

**Table 3.** Thermodynamic parameters for the adsorption of F– by GO/ES adsorbent

<table>
<thead>
<tr>
<th>Temperature (K)</th>
<th>K0</th>
<th>ΔG° (kJ/mol)</th>
<th>ΔS° (J/mol.K)</th>
<th>ΔH° (kJ/mol)</th>
</tr>
</thead>
<tbody>
<tr>
<td>298.15</td>
<td>0.9275</td>
<td>0.1865</td>
<td></td>
<td>-43.04</td>
</tr>
<tr>
<td>318.15</td>
<td>0.7230</td>
<td>0.8581</td>
<td></td>
<td>-12.73</td>
</tr>
<tr>
<td>333.15</td>
<td>0.5903</td>
<td>1.4560</td>
<td></td>
<td></td>
</tr>
<tr>
<td>348.15</td>
<td>0.4351</td>
<td>2.4085</td>
<td></td>
<td></td>
</tr>
</tbody>
</table>
adsorption of F⁻ unfavorable, while the negative value of
ΔS° indicates that the process is enthalpy driven and the
negative value of ΔH° implies that the adsorption
phenomenon is exothermic [12-13,19]. This explains that
the adsorption of F⁻ by using GO/ES adsorbent is not
suitable to be conducted at high temperature.

■ CONCLUSIONS

In this work, GO/ES adsorbent has been synthesized
for F⁻ removal application. The maximum adsorption
capacity of F⁻ (56 mg/g) was obtained at 25 °C, 25 mg/L of
F⁻ initial concentration, 0.05 g of GO/ES adsorbent dosage
and 120 min of contact time. Adsorption isotherms study
show that the experimental data were best fitted with
Langmuir isotherm, which indicates that F⁻ adsorption is
following monolayer adsorption on a homogenous
surface of GO/ES, with an R² value of 0.9983. The
adsorption kinetics of F⁻ by using GO/ES is best presented
by the pseudo-second-order model, involving a
chemisorption process. From the thermodynamic
analysis, F⁻ adsorption by using GO/ES adsorbent is a
spontaneous and exothermic process; hence, a higher
temperature is not favorable for this process.

■ ACKNOWLEDGMENTS

The authors are grateful for the financial support
and facilities provided by Universiti Teknologi MARA
(UiTM) to carry out this research work.

■ REFERENCES

[1] Mukherjee, S., and Halder, G., 2018, A review on the
sorptive elimination of fluoride from contaminated
short review: Removal of fluoride ions from ground
water by using various techniques, IJRG, 5 (4), 98–104.
[3] Nigri, E.M., Cechinel, M.A., Mayer, D.A., Mazur,
L.P., Loureiro, J.M., Rocha, S.D., and Vilar, V.J.,
2017, Cow bones char as a green sorbent for fluorides
removal from aqueous solutions: Batch and fixed-bed
standard for drinking water quality, Ministry of Health
Malaysia.
[5] Department of Environment Malaysia, 2010,
Environmental requirements: A guide for investors,
11th Ed., Ministry of Natural Resources and
Environment.
groundwater by adsorption technology - The
occurrence, adsorbent synthesis, regeneration and
disposal, Dissertation, Delft University of
Technology, Delft, Netherlands.
and Madaree, A.M., 2015, Removal of fluoride from
urban drinking water by eggshell powder, Bulg.
2015, Synthesis of graphene oxide and its
application for efficient removal of fluoride from
272.
Graphene oxide and its application as an adsorbent
Biotechnol., 89 (2), 196–205.
[10] Narasimharao, K., Venkata, R.G., Sreedhar, D., and
Vasudevarao, V., 2016, Synthesis of graphene oxide
by modified Hummers method and hydrothermal
synthesis of graphene-NiO nano composite for
supercapacitor application, J. Mater. Sci. Eng., 5 (6),
1000284.
[11] Liu, L., Cui, Z., Ma, Q., Cui, W., and Zhang, X.,
2016, One-step synthesis of magnetic iron-
aluminum oxide/graphene oxide nanoparticles as a
selective adsorbent for fluoride removal from
aqueous solution, RSC Adv., 6 (13), 10783–10791.
[12] Mohammad-Rezaei, R., Razmi, H., and Dehgan-
Reyhan, S., 2014, Preparation of graphene oxide
doped eggshell membrane bioplatform modified
Prussian blue nanoparticles as a sensitive hydrogen
K.C., Das, C., Baneerjee, A., and Datta, J.K., 2012,
Eggshell powder as an adsorbent for removal of
fluoride from aqueous solution: Equilibrium,


